Simulation Study of the Formation of Corrosive Gases in Coal Combustion in an Entrained Flow Reactor

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Abstract: Gaseous sulfur species play a major role in high temperature corrosion of pulverized coal fired furnaces. The prediction of sulfur species concentrations by 3D-Computational Fluid Dynamics (CFD) simulation allows the identification of furnace wall regions that are exposed to corrosive gases, so that countermeasures against corrosion can be applied. In the present work, a model for the release of sulfur and chlorine species during coal combustion is presented. The model is based on the mineral matter transformation of sulfur and chlorine bearing minerals under coal combustion conditions. The model is appended to a detailed reaction mechanism for gaseous sulfur and chlorine species and hydrocarbon related reactions, as well as a global three-step mechanism for coal devolatilization, char combustion, and char gasification. Experiments in an entrained flow were carried out to validate the developed model. Three-dimensional numerical simulations of an entrained flow reactor were performed by CFD using the developed model. Calculated concentrations of SO2, H2S, COS, and HCl showed good agreement with the measurements. Hence, the developed model can be regarded as a reliable method for the prediction of corrosive sulfur and chlorine species in coal fired furnaces. Further improvement is needed in the prediction of some minor trace species.

Keywords: Computational Fluid Dynamics (CFD); high temperature corrosion; pulverized fuel-firing systems

1. Introduction

Coal is one of the most abundant energy sources and accounts for up to one-third of the world’s primary energy consumption in 2018 [1], and forecasts see only a slight decrease in coal consumption over the next twenty years [2]. During the last few decades, major efforts have been made to minimize pollutant emissions from coal-fired power plants. Air staging as a primary measure to reduce NOx emissions is generally applied in furnaces. However, the reducing conditions prevailing in the air staging area cause the formation of corrosive sulfur species, such as H2S and COS, which can lead to high-temperature corrosion on the furnace walls [3,4].

Despite their importance, the reaction kinetics causing the formation of these sulfur species are still not fully understood. The complex kinetics, which comprise a large number of species and different kinds of elementary reactions, are difficult to determine experimentally. To describe the sulfur reactions in detail, elaborate reaction mechanisms are necessary, which are not applicable to be linked to Computational Fluid Dynamics (CFD) Simulations. However, the computing time of CFD simulations with several million grid cells becomes manageable when reduced reaction mechanisms are used. The mechanism has to be chosen carefully regarding the operation conditions of the investigated facility.
The procedure of combining experiments in a drop tube furnace and CFD simulations is often used to gain information about combustion processes, develop and validate reaction models, or investigate pollutant formation [5–9]. One can find many studies on simulation of NO\textsubscript{x} emissions in the literature. Han et al. [10] implemented a skeletal mechanism into a CFD code and carried out simulations of an entrained flow reactor to predict NO\textsubscript{x} emissions. Wei et al. [11] added sulfur chemistry to the reaction mechanism. They predicted CO, NO\textsubscript{x}, and SO\textsubscript{x} concentration profiles in the entrained flow reactor by CFD simulations. Lee et al. [12] investigated in a drop tube furnace the influence of coal volatile matter content on fuel NO emissions. They performed CFD simulations with a simple reaction model and obtained a good agreement between measurements and calculations. Other studies focus on the prediction of SO\textsubscript{x} emissions with simplified models and mechanisms [13,14]. Müller et al. [15] include several release mechanisms for sulfur in their simulation, but, for gas phase chemistry, only global reactions were used. This method shows good results for SO\textsubscript{x} emission prediction. However, the formation of intermediate species, like COS, is not possible. Maffei et al. [16] developed a comprehensive model of sulfur release from coal. It includes 15 species and 15 reactions and was validated against literature data. But they performed only 0D calculations and no CFD simulations.

There are few studies regarding prediction of corrosive atmospheres by numerical simulation. Modlinski and Hardy [17] implemented four global reaction mechanisms into a CFD code to predict O\textsubscript{2} and CO concentrations in a power plant furnace. They proposed an on-line corrosion risk monitoring system based on O\textsubscript{2} and CO measurements and tried to adjust the demonstrated monitoring system based on the identification of reducing atmospheres. They do not incorporate trace species, like H\textsubscript{2}S or COS. Most researchers investigate fly-ash deposition on furnace walls and heat exchangers. Li et al. [18] give a comprehensive overview of CFD-based prediction of fly-ash deposition in their review paper. Nevertheless, they do not consider corrosive atmospheres. Wang et al. [19] proposed a numerical method to predict the sulfuric acid dew point temperature. With this information, they assess the corrosion risk due to sulfuric acid dew point corrosion. They do not calculate any species concentrations.

In this study, a reaction model was developed that strikes a balance between accuracy and computational effort of CFD simulations with large computational grids. The model is incorporated in 3D-CFD simulations of an entrained flow reactor. In the presented work, not only are major pollutants, like CO and SO\textsubscript{2}, predicted by CFD simulations but also the concentrations of trace species, like COS and H\textsubscript{2}S. Based on studies in the literature [20] and previous work [21], this article postulates an extensive reaction mechanism to describe the formation of sulfur species and their reaction with other non-sulfur species, e.g., hydrocarbons, oxygen, and carbon-oxides, during coal combustion. The existing model, which comprised the gas phase mechanism, was updated to include sulfur species release during combustion. Three-dimensional numerical simulations of coal combustion in an entrained flow reactor at different stoichiometric ratios and temperatures were performed. The sulfur emissions of a U.S. high volatile C bituminous coal were studied by combustion experiments in the entrained flow reactor. Four different air ratios (0.7, 0.8, 0.9, and 1.0) and three different temperatures (1000 °C, 1150 °C, and 1300 °C) were investigated. To validate the model, the simulation results are compared to measurements conducted in the entrained flow reactor.

2. Reaction Model

The objective of the simulation is the prediction of sulfur containing compounds’ behavior during different stages of the coal combustion process by numerical modeling. The sulfur chemistry applicable during coal combustion, which forms the basis of the model, was already described in an earlier work from us [21] and was expanded to describe the release processes of sulfur in more detail. Mineral matter transformation of sulfur bearing minerals was added to the model. The modeling approach is described extensively in a previous article [22] and will be summarized in the following.
2.1. Reaction Mechanism of Solid State Reactions

Depending on type, rank and grade of the coal the sulfur content ranges from 0.3% to 5.0%. In addition to its elemental form, which is found only in trace amounts, sulfur in coal can occur in two different forms. Inorganic and organic sulfur are the major types, and the distribution between them strongly depends on the considered coal [23]; in some cases, up to 40% of inorganic sulfur have been found in coal. In case the sulfur atoms are bound directly in the carbon chains, it is termed organic sulfur. Various functional groups have been found in analyses, e.g., thiophenic and thiolic sulfur, aliphatic sulfides, and aromatic disulfides [16,24], accounting for a high complexity. The other major sulfur source in coal is inorganic sulfur, where the sulfur atoms occur in minerals. Pyrite comprises the majority of the inorganic sulfur in coal. Depending on the origin of the coal, the amount of sulfates can also be high. Gypsum is, among others, the main source of sulfates [25,26]. Hydrogen sulfide is mainly formed from organic sulfur during pyrolysis [27,28], while inorganic sulfur is released during the transformation of mineral matter. Inorganic sulfur is accounted for as pyritic sulfur that occurs as pyrite and sulfates, which may occur as gypsum and other compounds. As gypsum represents the majority of the sulfates [26], no further sulfates are included in the model. The numerical costs of expanding the model by additional sulfate compounds would exceed the benefit of a more complex model. Organic bound sulfur is not further subdivided.

In this section, the mineral matter transformation reactions are characterized. Twelve reactions are included in the model, which are describing the fate of pyrite and gypsum in oxidizing and reducing conditions, as well as the evaporation of sodium chloride. The reaction mechanism is given in Table 1. The oxidation of pyrite starts at temperatures of approximately 330°C, leading to the formation SO$_2$ (c.f. reaction 3, Reference [29]). In inert atmosphere, meaning absence of oxygen, and at temperatures above 570 °C, pyrite decomposes to pyrrhotite (c.f. reaction 1, Reference [25]), which then decomposes into iron and sulfur above 730 °C (c.f. reaction 4). In the presence of oxygen iron is oxidized to iron oxides. During this reaction, hematite (Fe$_2$O$_3$) or magnetite (Fe$_3$O$_4$) can be formed as products. As the main product magnetite was found during pyrite combustion in drop tube experiments [30], so, for convenience, only magnetite is included in the model. It shall be mentioned that the oxidation of iron is not included in the model to reduce the computational effort. This particular reaction is not important for the prediction of corrosive atmospheres. In oxygen rich atmosphere, Pyrrhotite is also oxidized above 820 °C to form sulfur dioxide and magnetite according to reaction 2. Gypsum is decomposed in several steps. First, the gypsum is dehydrated according to reaction 5 [29], forming water and anhydrite. This dehydration can proceed already at low temperatures of 60 °C. The second step starts at a temperature approximately 830 °C with the decomposition of anhydrite to calcium oxide and sulfur trioxide, which reacts to sulfur dioxide and oxygen at high temperatures (c.f. reaction 6). Sulfur trioxide is neglected in the reaction mechanism, since the description of the reactions in high temperature areas of furnaces is the overall aim of the model. In oxygen rich areas, reaction 6 is promoted. In case of the absence of oxygen, anhydrite reacts mainly to calcium sulfide and carbon dioxide according to reaction 7. According to reaction 9 calcium sulfide is oxidized to calcium oxide and sulfur dioxide [31]. Water vapor can, in the absence of oxygen, serve as an alternative oxidant, especially at high temperatures (c.f. reaction 10). In this case, the gaseous product is hydrogen sulfide. Sulfur dioxide was also proven in experiments as a product of reaction 10 [32]. However, this reaction pathway is neglected, since in the presence of oxygen hydrogen sulfide is anyway rapidly oxidized to sulfur dioxide [21]. The share of NaCl in the total chlorine content of coal is generally between 70–80% [33,34]. The sodium chloride evaporates quickly during the pyrolysis of the coal and is then present in gaseous form (c.f. reaction 11). The gaseous sodium chloride can be bound again by reactions with aluminum silicates in the fly ash. The products of this reaction are nepheline and hydrogen chloride (c.f. reaction 12) [35].
2.2. Kinetics of the Solid State Reactions

Information about individual reaction steps in solid state kinetics is difficult to obtain. In order to draw mechanistic interpretations, it is necessary to identify an suitable reaction model. One model is generally appropriate to describe a particular reaction category and provide a rate equation, but different reaction types are required for different model approaches. The reaction rates of the solid state reactions can be evaluated generally with the following kinetic equation [36]:

\[
\frac{d\alpha}{dt} = K \cdot f(\alpha).
\]  (1)

Here, \(\alpha\) represents the conversion fraction, \(K\) describes the rate constant and \(f(\alpha)\) is the kinetic model function, which is used to describe the specific reaction process. The conversion fraction is calculated on the basis of the mass fraction and is defined as [37]:

\[
\alpha = \frac{m_0 - m_t}{m_0 - m_{final}}.
\]  (2)

\(m_0\) describes the mass at the beginning, \(m_t\) the mass at time \(t\), and \(m_{final}\) the remaining mass after the reaction. In this study, a total conversion is assumed, this means \(m_{final} = 0\). The model function \(f(\alpha)\) defines the reaction type, i.e., the physical or chemical process. Several basic functions for idealized models of solid state reactions are available in the literature [36,37]. The functions used in this study are given in Table 2. These are applicable to reactions which are controlled by only one physical process or chemical reaction. The formal kinetic nth-order model (c.f. function (1) in Table 2) is used for the heterogeneous reactions of mineral matter transformation [38]. The model order is set to 1. For the dehydration process of gypsum, in contrast to all other reactions, the reaction of calcium sulfide with oxygen (c.f. reaction 8) is described using a Jander equation (c.f. function (3)) as a mechanism function [31]. The authors obtained the best results for the reaction of CaS with Oxygen with this mechanism function. The evaporation of NaCl is modeled with an Arrhenius approach. The kinetic parameters are listed in Table 1. The secondary binding of NaCl into the fly ash is modeled as a function of the particle surface \(A_p\) and the mass of Metakaolinite \(m_{kao}\):

\[
\frac{dm_{kao}}{dt} = -k_{eff} \cdot A_p \cdot m_{kao}.
\]  (3)

### Table 1. Reactions of mineral matter transformation.

<table>
<thead>
<tr>
<th>No.</th>
<th>Reaction</th>
<th>(E) [kJ/mol]</th>
<th>(A)</th>
</tr>
</thead>
<tbody>
<tr>
<td>1</td>
<td>(FeS_2(s) \rightarrow \frac{1}{2}Fe_2S_3(s) + \frac{3x}{4}S_2)</td>
<td>92.04</td>
<td>195.9 1/s</td>
</tr>
<tr>
<td>2</td>
<td>(Fe_2S_3(s) + 2\frac{x}{2}O_2 \rightarrow \frac{x}{2}Fe_3O_4(s) + SO_2)</td>
<td>92.5</td>
<td>(2.1 \times 10^{-5}) 1/sPa</td>
</tr>
<tr>
<td>3</td>
<td>(FeS_2(s) + 8O_2 \rightarrow Fe_3O_4(s) + 6SO_2)</td>
<td>87.9</td>
<td>0.017 1/sPa</td>
</tr>
<tr>
<td>4</td>
<td>(FeS_3(s) \rightarrow xFe(s) + 0.5S_2)</td>
<td>92.5</td>
<td>0.45 1/s</td>
</tr>
<tr>
<td>5</td>
<td>(CaSO_4 \times 2H_2O(s) \rightarrow CaSO_4(s) + 2H_2O)</td>
<td>120</td>
<td>(3.2 \times 10^{15}) 1/s</td>
</tr>
<tr>
<td>6</td>
<td>(CaSO_4(s) \rightarrow CaO(s) + SO_2 + 0.5O_2)</td>
<td>506.2</td>
<td>(5.24 \times 10^{14}) 1/s</td>
</tr>
<tr>
<td>7</td>
<td>(CaSO_4(s) + 4CO \rightarrow CaS(s) + 4CO_2)</td>
<td>390</td>
<td>(2.2 \times 10^{14}) 1/s</td>
</tr>
<tr>
<td>8</td>
<td>(CaSO_4(s) + 4CO \rightarrow CaO(s) + 3CO_2 + COS)</td>
<td>280.5</td>
<td>(1.01 \times 10^{13}) 1/s</td>
</tr>
<tr>
<td>9</td>
<td>(CaS(s) + 1.5O_2 \rightarrow CaO(s) + SO_2)</td>
<td>357.4</td>
<td>(1.04 \times 10^{13}) 1/s</td>
</tr>
<tr>
<td>10</td>
<td>(CaS(s) + H_2O \rightarrow CaO(s) + H_2S)</td>
<td>115.4</td>
<td>481.95 1/s</td>
</tr>
<tr>
<td>11</td>
<td>(NaCl(s) \rightarrow NaCl(g))</td>
<td>239</td>
<td>(8.2 \times 10^8) 1/s</td>
</tr>
<tr>
<td>12</td>
<td>(2NaCl(g) + H_2O + Al_2O_3(s) \times 2SiO_2 \rightarrow 2NaAlSiO_4(s) + 2HCl(g))</td>
<td>41.5</td>
<td>2.5 kg/m^2sPa</td>
</tr>
</tbody>
</table>
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Table 2. Mechanism functions of the solid state reactions.

<table>
<thead>
<tr>
<th>No.</th>
<th>Function</th>
</tr>
</thead>
<tbody>
<tr>
<td>1</td>
<td>$f(\alpha) = (1 - \alpha)^n$</td>
</tr>
<tr>
<td>2</td>
<td>$f(\alpha) = 2(1 - \alpha)[-\ln(1 - \alpha)]^{0.5}$</td>
</tr>
<tr>
<td>3</td>
<td>$f(\alpha) = 1.5(1 - \alpha)^{2/3}[1 - (1 - \alpha)^{1/3}]^{-1}$</td>
</tr>
</tbody>
</table>

The reaction rate is limited on the one hand by the diffusion of NaCl into the particles, expressed by $k_D$, and on the other hand by the chemical reaction rate $k_{ch}$. Another influencing factor is the partial pressure of NaCl, $p_{NaCl}$. The effective reaction rate $k_{eff}$ is then obtained:

$$k_{eff} = \frac{k_{ch} \cdot k_D}{k_{ch} + k_D} \cdot p_{NaCl}. \quad (4)$$

The diffusion rate of sodium chloride is determined according to Equation (5):

$$k_D = \frac{2M_{meta} \cdot D_{NaCl}}{\alpha_p \cdot R \cdot T_p}. \quad (5)$$

The diffusion coefficient of NaCl $D_{NaCl}$ is $1425 \times 10^{-10} \cdot T^{1.88}$ [34].

The rate constant $K$ in Equation (1), as well as $k_{chem}$ in Equation (4), is computed by an Arrhenius approach. The kinetic parameters are obtained from isothermal kinetic data, which are given in literature for every particular reaction. The parameters for each reaction are listed in Table 1.

3. Experimental Setup

For validation, sulfur release during coal combustion was investigated in an entrained flow reactor. Details of the reactor design are given in Figure 1.

Gas and particles were sampled in the reaction zone. The concentrations of sulfur species SO$_2$, H$_2$S, and COS were measured by a mass spectrometer. The coal used in this study was a U.S. high volatile C bituminous coal with high sulfur content. Disulfide, sulfate, and organic sulfur make up approximately 15%, 16%, and 69% of total sulfur, respectively, while sulfides were not present in the coal. Additional properties of the coal are given in Table 3. A detailed description of how coal characteristics (proximate and ultimate analysis and sulfur species) were obtained is given in Reference [39].

Table 3. Coal analysis in wt.-%, dry basis.

<table>
<thead>
<tr>
<th>Proximate Analysis</th>
<th>Ultimate Analysis</th>
</tr>
</thead>
<tbody>
<tr>
<td>Ash</td>
<td>9.33</td>
</tr>
<tr>
<td>Volatiles</td>
<td>37.3</td>
</tr>
<tr>
<td>Fixed Carbon</td>
<td>53.37</td>
</tr>
<tr>
<td>C</td>
<td>72.91</td>
</tr>
<tr>
<td>H</td>
<td>4.97</td>
</tr>
<tr>
<td>N</td>
<td>1.25</td>
</tr>
<tr>
<td>S</td>
<td>3.01</td>
</tr>
</tbody>
</table>

The main component of the entrained flow reactor is a vertical reaction zone made of ceramic with a total length of 2200 mm and an ID of 70 mm. Electrical heating elements allow the adjustment of wall temperature of the reaction zone between 900 °C and 1600 °C. The fuel feed to the reaction zone is located on the top and consists of an injection nozzle connected to a water-cooled lance. Coal particles are gravimetrically dispensed from a container by a screw conveyor, transported through the water-cooled injection lance using a carrier gas stream, and enter the reaction zone. A second gas stream (co-flow) is led through the preheating section of the entrained flow reactor where it is heated up to a specified temperature, usually the wall temperature of the reaction zone. The co-flow enters the reaction zone through an annual gap located on the top, thus enabling an immediate mixing with
the cold particle-loaded carrier gas stream to achieve high heating rates of the particles. Particles and gas move downward through the reaction zone at approximately 1 m/s. The resulting product gas is sampled using a ceramic tube entering the reaction zone perpendicularly through a bore. Sixty liters per hour or 3 to 5% of total gas flow are suctioned from the reaction zone. Particles are removed from the sampling gas stream directly at the outlet of the sampling port by a fine filter made of PTFE. The sampling gas is led to the gas analysis device in flexible PTFE tubes. Filter and tubes are heated to 180 °C to prevent the condensation of water throughout the sampling line. A quadrupole mass spectrometer (InProcess Instruments GAM 200) with a mass range of 100 a.m.u. is used for gas analysis. It is calibrated to determine the concentration of 11 species in the sampling gas, including SO₂, H₂S, and COS. Product gas and remaining particles (i.e., ash) leave the reaction zone on the bottom and are quenched by water injection.

All validation experiments were conducted following the same general scheme. The reactor was heated to the specified temperature (1000 °C, 1150 °C, or 1300 °C, respectively). An air flow corresponding to the desired stoichiometric ratio was introduced to the reactor and maintained by mass flow controllers. A coal mass flow of 0.2 kg/h was then fed in the carrier gas stream. Product gas and particles were sampled at two measuring planes, 1300 mm and 2000 mm downstream of the
injection nozzle, corresponding to residence times of approximately 1 s and 2 s. When stable conditions were achieved (i.e., the concentrations measured by gas analysis reach a steady state), the operating point was maintained for approximately 20 min, and concentrations were recorded.

4. Numerical Setup

The introduced model of mineral matter conversion was combined with the existing model. The model describes coal combustion and formation of sulfur species [21]. The detailed reaction mechanism of the gas phase sulfur chemistry, which is based on the sulfur mechanism of the University of Leeds [20], was updated. A reduced chemical mechanism is used to describe the C-H-O combustion system. It contains 73 elementary reactions among 17 species [40]. Reactions of chlorine species were added based on the mechanism of Pelucchi et al. [41]. The complete reaction mechanism is given in the Appendix A. It contains 48 species and 152 elementary reactions.

Numerical simulations of coal combustion in the entrained flow reactor were performed to validate the developed model. A 3D numerical grid with 505,850 computational cells was used to discretize the geometry of the reaction zone, as illustrated in Figure 2. The dimensions of the computational domain are 2200 mm by 70 mm. The grid at the fuel inlet is very fine in order to reach a high spatial resolution of 0.8 mm in the combustion zone and to map the geometry of the injection nozzle. Towards the outlet, grid size decreases. Only the upper part of the reaction tube is shown in Figure 2 for a clearer presentation because the diameter of the tube is much smaller than the height. Reynolds-averaged transport equations are solved using the CFD simulation software FLUENT for flow, energy, and species fields. The standard \( k-\epsilon \) model is used to account for turbulence. The eddy-dissipation concept [42] (EDC) allows including detailed chemical mechanisms in the simulation of turbulent flows and is therefore applied for turbulence–chemistry interaction. Radiative heat transfer is modeled with the discrete ordinates (DO) radiation model. Gas–solid interaction is considered for using the mixture model based on mixture theory. In this model solid particles are treated as a heavy gas component in the gas–solid mixture. Coal combustion is described with an enhanced version of an in-house coal combustion code, Eulerian Simulation Tool for Solid Fuels (ESTOS). This is integrated into FLUENT through user-defined functions. The standard version of ESTOS is based on a five-step global reaction scheme for coal combustion: Coal pyrolysis, char oxidation, char gasification with CO\(_2\), combustion of hydrocarbons, and CO oxidation and has been validated comprehensively against measurements in lab-scale furnaces and full-scale boilers [43–45]. In the extended version of ESTOS, the two global gas-phase reactions are substituted by the detailed gas-phase kinetics mentioned above. The transformation of mineral matter is included to the coal combustion and gasification model. Four particle size classes of the coal particles are included in the simulation (c.f. Table 4). A sieve analysis has been carried out and based on the results four equivalent size classes have been calculated.

<table>
<thead>
<tr>
<th>Size Class 1</th>
<th>Size Class 2</th>
<th>Size Class 3</th>
<th>Size Class 4</th>
</tr>
</thead>
<tbody>
<tr>
<td>76 ( \mu )m</td>
<td>87 ( \mu )m</td>
<td>110 ( \mu )m</td>
<td>167 ( \mu )m</td>
</tr>
</tbody>
</table>
5. Results and Discussion

Overall, twelve combinations of wall temperature and air ratio have been investigated. Three different wall temperatures (1000 °C, 1150 °C, and 1300 °C) were applied, and, for each wall temperature, four air ratios from 0.7 to 1.0 were considered. Reducing conditions were created to have corrosive gases formed in the reaction tube. Gas measurements were conducted at two positions, 1300 mm (measuring plane 1) and 2000 mm (measuring plane 2) downstream of the injection nozzle. The total length of the reaction tube is 2200 mm. So, measuring plane 1 is just after half the length of the tube, and measuring plane 2 is almost at the end of the tube. Measurements were conducted at two different residence times of the particles in the reaction tube. At measuring plane 1, the residence time is about 0.5 s and at measuring plane 2, the residence time is between 1 s and 1.5 s, also depending on the applied boundary conditions [46].
5.1. Temperature and Velocity

The calculated 3D distributions of temperature and velocity in the upper part (0–800 mm axial distance) of the reaction tube for air ratio of 1.0, and three wall temperatures are shown in Figures 3 and 4.

The coal dust and transport gas enters the reaction tube through the injection nozzle with low temperature of 20 °C. The secondary air flows through the outer annulus (c.f. co-flow in Figure 2) into the reaction tube and is preheated to the employed wall temperature.

For wall temperature of 1300 °C, the highest temperature of 1800 °C is reached in the combustion zone. The temperature level is lower for the two other cases. Downstream the combustion zone, the gas temperature reaches the wall temperature. After an axial distance of 350 mm, the temperature remains constant, indicating that the conversion of volatiles and char is completed. At a wall temperature of 1300 °C, gas temperature declines over a longer distance after the burnout zone until it matches the wall temperature. Homogeneous temperature distribution is reached on both measuring planes.

![Figure 3. Temperature in the first 800 mm of the reaction tube for the three different wall temperature cases with $\lambda = 1.0$.](image)

Figure 4 shows the axial velocity in the reaction tube for the three temperature cases at an air ratio of 1.0. There is a velocity peak at the exit of the injection nozzle in all cases. The peak value reaches 1.4 m/s at wall temperature 1300 °C, 1.2 m/s at wall temperature 1150 °C, and 1.1 m/s at wall temperature 1000 °C. The co-flow velocity ranges from 0.45 m/s to 0.35 m/s. Shortly after the
combustion zone, a laminar flow occurs. The velocity in the middle of the tube correlates with the wall temperature. The higher the temperature, the higher is the velocity in the tube (c.f. Figure 4).

![Figure 4. Z Velocity in the first 800 mm of the reaction tube for the three different temperature cases with $\lambda = 1.0$.](image)

5.2. Main Species

The simulation results are first evaluated with regard to the main species concentrations.

The concentrations of oxygen and carbon monoxide are shown in Figure 5 for all four air ratios at wall temperature of 1300 °C. The contour plots illustrate the oxygen concentration on the left half of the reaction tube and the carbon monoxide concentration on the right half. There are two different legends for the volume fractions at the top of the figure. Coal and air enter the reaction zone from the top through the injection lance in the middle, in which the walls are shown as a white space in the plots. Oxygen is quickly consumed by volatiles and char while CO is produced, leading to the temperature peak close to the inlet. There is a high CO concentration of 16 Vol-% in the reaction tube for air ratio 0.7 because there is not enough oxygen to oxide the CO to CO$_2$. The remaining CO concentration is lower for the other air ratios; it amounts to 11 Vol-% for air ratio 0.8, 5.7 Vol-% for air ratio 0.9, and almost zero for air ratio of 1.0. The oxygen is rapidly consumed and the oxygen concentration reaches zero fast, although it takes slightly longer with increasing air ratio. The conditions in the reaction tube are very similar for the other wall temperatures. The combustion process is just slower, and the peak temperatures are lower, as it can be seen in Figure 3.
Figure 5. Oxygen and carbon monoxide concentrations for the four air ratios at wall temperature 1300 °C in the upper 800 mm of the reaction tube. The oxygen concentration is given on the left side, and the carbon monoxide concentration on the right side, for each air ratio in the assembled contours.

The calculated main species concentrations are compared to measurements to evaluate the accuracy of the prediction of the overall combustion process. Especially, the CO and O₂ concentrations are important boundary conditions when calculating the concentrations of trace species. Trace species like H₂S, COS, and HCl are only formed in reducing atmosphere and they are oxidized very fast. All concentration values are given on dry basis in the following discussion. Error bars in the diagrams
indicate a measurement uncertainty. The uncertainty of the measured species concentrations is estimated by a consideration of possible error sources and the used equipment. While the gas flows were controlled within a narrow tolerance range, fluctuations in the coal feed rate may occur. Fluctuations in the coal feed rate directly affected the ratio of fuel to oxygen and therefore the measured gas concentrations. For the experiments, a conservative estimate is ±2% for the CO₂ concentration and CO, ±100 ppm for the total sulfur concentration and ±10 ppm for total chlorine species concentration. At each measurement point, at least 30 cycles of the mass spectrometer’s measuring routine were recorded. Average concentration values are calculated from the measured sequences. In all concentration measurement results shown, error bars indicate the standard deviation of the recorded measurement data.

Figure 6 shows the concentrations of CO₂ and CO for an air ratio of 1.0. The blue symbols indicate the measurements, and the red symbols indicate the simulation results. There is a good agreement in CO₂ concentrations for all temperatures at both measuring planes. The CO₂ concentration is 16.6 Vol-% for all wall temperatures. No significant difference in the CO₂ concentration between the different wall temperatures indicates a complete burnout of the fuel also for the lower temperatures at air ratio 1.0. The calculated CO concentrations are almost zero for all temperatures at both measuring planes. The measured CO concentrations are between zero and 4.5 Vol-%. For wall temperature 1000 °C, there are 3.8 Vol-% CO at measuring plane 1, and the CO concentration decreases downstream the reaction tube like it is predicted in the simulation. For wall temperature 1150 °C, the measured CO concentration stays at about 4.4 Vol-%. This is not reasonable because there is enough oxygen to oxidize all of the CO at air ratio of 1.0. At wall temperature 1300 °C, there is no CO measured at measuring plane 1 and 4.5 Vol-% at lane 2. The elevated CO concentrations are not present in the simulation results. They can be an indication of a delayed combustion. In this case, the residence time is not high enough to ensure a complete burnout. The measured CO concentrations are in contradiction to the operation conditions. The concentration should be small in case of good mixing and a complete combustion, like it is the case in the simulations.

Figure 7 shows the CO₂ and CO concentrations for air ratio 0.9 over the three wall temperatures. The CO concentrations are now elevated up to 6 Vol-%. For measuring plane 1, only one CO measurement is available, namely for wall temperature 1150 °C. The simulation results show the same CO concentration of 5.6 Vol-%. The measured CO concentration for wall temperature 1300 °C at measuring plane 2 is with 2.5 Vol-% unusually low and out of line. The calculated CO concentration is much higher with 5.6 Vol-%. Over the three wall temperatures, the CO concentrations show a slightly increasing trend at air ratio 0.9. The measured and calculated CO₂ concentrations are lower than at
air ratio 1.0. The calculated CO$_2$ concentrations are lower than the measurements, except for wall
temperature of 1000 °C at measuring plane 1. Nevertheless, all the other calculated concentrations
are within the error bars of the measurements. Two measured CO$_2$ concentrations are clearly out of
line, namely at wall temperature 1300 °C at measuring plane 1 and at wall temperature 1000 °C at
measuring plane 2. With 20 Vol-%, they are higher than at air ratio 1.0. The reason could be an elevated
coal mass flow during the measurements or measurement inaccuracy. Despite this, the major part of
the measurements and simulation results are in good agreement for the air ratio of 0.9.

![Image of CO$_2$ and CO concentrations for air ratio 0.9](image)

**Figure 7.** Measured and simulated CO$_2$ and CO concentrations for air ratio 0.9.

To have a closer look at the effect of a lower air ratio, Figures 8 and 9 show the CO and the
CO$_2$ concentrations for all four air ratios for wall temperature of 1300 °C. The CO concentrations
(c.f. Figure 8) show a decreasing trend from 15 Vol-% to 0 Vol-% with increasing air ratio in the
simulations. The measured concentration for air ratio 0.9 is higher than for air ratio 0.8. It is believed
that the CO measurement does not represent the real condition for air ratio 0.9. The simulation results
show a clear linear behavior because the oxygen supply is reduced with decreasing air ratio, and the
coal mass flow was kept constant. Almost all calculated concentrations are within the error bars of
the measurements.

![Image of CO concentrations for different air ratios](image)

**Figure 8.** Measured and simulated CO concentrations for wall temperature 1300 °C.

The CO$_2$ concentrations show a opposite behavior compared to the CO concentrations. With increasing air ratios, the CO$_2$ concentrations increase from 5 Vol-% to 16.6 Vol-% in the simulations.
The measurements show the same trend, except for air ratio 0.7. At measuring plane 2, the measured CO\textsubscript{2} concentrations is with 12.5 Vol-% even higher than at air ratio 0.8. The measured CO\textsubscript{2} concentration is also elevated at measuring plane 1 for air ratio 0.7. A temporary higher coal feed rate may be the reason for this increased concentrations. For the other three air ratios, there is a very good agreement between measurements and simulation results at both measuring planes.

![Graph a) Measuring plane 1](https://example.com/graph_a)

![Graph b) Measuring plane 2](https://example.com/graph_b)

**Figure 9.** Measured and simulated CO\textsubscript{2} concentrations for wall temperature 1300 °C.

A reason for the high CO concentrations at some operation points, which seem to be out of line (e.g., air ratio 1.0 c.f. Figure 6 or air ratio 0.9 in Figure 8), despite a sufficient oxygen amount could be insufficient mixing in the reaction tube. A fully developed pipe flow is achieved in the tube downstream the injection nozzle. The mass transfer in radial direction seems to be too small to allow a complete burnout. Hence not all CO is oxidized. The overall agreement between simulation results and measurement is satisfying. The boundary conditions for the trace species calculations are provided sufficiently accurate by the reaction model.

### 5.3. Trace Species

Sulfur and chlorine species, like H\textsubscript{2}S, COS, and HCl, play a major role in high temperature corrosion processes [3,4]. They occur only in small amounts off parts per million (ppm) in coal fired furnaces. The simulation results are discussed regarding the concentrations of SO\textsubscript{2}, H\textsubscript{2}S, COS, and HCl in this section. Figure 10 shows the H\textsubscript{2}S and SO\textsubscript{2} concentrations for wall temperature 1300 °C for all four air ratios.

![Graph for H\textsubscript{2}S and SO\textsubscript{2}](https://example.com/graph)

**Figure 10.** Measured and simulated H\textsubscript{2}S and SO\textsubscript{2} concentrations for wall temperature 1300 °C at measuring plane 2.
H$_2$S concentrations increase with decreasing air ratio. SO$_2$ concentrations show an opposite trend. The H$_2$S concentrations rise in the reducing atmospheres along with the CO concentrations, like it is depicted in Figure 8. With lack of oxygen, H$_2$S cannot be oxidized, and SO$_2$ is in addition converted to H$_2$S under reducing conditions. Both, measurements and simulation results, show a higher SO$_2$ concentration for air ratio 0.9 than for air ratio 1.0. The sulfur release in form of SO$_2$ is elevated. In the reducing atmosphere, the SO$_2$ is reduced to H$_2$S. This effect is under predicted in the simulation for a high air ratio and the calculated SO$_2$ concentration is higher than the measurement. Nevertheless, for air ratio 0.8, the results are in good agreement with the measurements. For air ratio 0.7, almost no SO$_2$ is present in the simulation. The measurements show a concentration of 350 ppm SO$_2$ for air ratio 0.7, which is a higher concentration than for air ratio 0.8. The result is still acceptable because the calculated SO$_2$ concentration is within the error margin of the measurement. The H$_2$S concentration rises up to 2000 ppm for air ratio 0.7 and declines to zero for air ratio 1.0 in the simulation. The measurements show 400 ppm H$_2$S for air ratio 1.0. The high concentration can not be explained with the operation conditions. Due to the fully developed pipe flow in the reaction tube, there might be bad mixing in the tube, and not all H$_2$S is oxidized. The same is observed for the CO concentrations at air ratio 1.0 (c.f. Figure 6) The simulation results and the measurements are in good agreement for all air ratios, whereby the three substoichiometric air ratios are of particular importance.

Figure 11 shows the H$_2$S and SO$_2$ concentrations for the air ratios 0.7, 0.8, and 0.9 for the three wall temperatures at measuring plane 2. The further discussion of the trace species will be limited to the substoichiometric air ratios 0.7, 0.8, and 0.9 because for air ratio 1.0 no significant amounts of the trace species occur. The H$_2$S concentrations show a peak for wall temperature 1150 °C. Lower amounts of H$_2$S are present for the other wall temperatures. For elevated temperatures, the sulfur release is higher than for lower temperatures. The reaction velocity is highly temperature dependent; therefore, at 1300 °C, the H$_2$S can react faster to other sulfur species. This results in lower H$_2$S concentrations at 1300 °C, although the sulfur release from the coal and minerals is high. With decreasing air ratio, the H$_2$S concentrations increase slightly for all temperatures. The simulation predicts the H$_2$S concentrations accurately. Just for wall temperature 1000° at air ratio 0.9, the calculated H$_2$S concentrations are clearly higher than the measurement. The simulation over predicts the sulfur release for this condition and the oxidation of H$_2$S to SO$_2$ is under predicted. This can also bee seen by the elevated SO$_2$ concentration in the measurements for the wall temperature of 1000° at air ratio 0.9. The calculated SO$_2$ concentrations is almost zero, clearly lower than the measurement. For wall temperature 1150° at air ratio 0.8, the calculated H$_2$S concentration is also higher than the measurement. At this point, the measured and calculated SO$_2$ concentrations are in good agreement. Either the sulfur release is overpredicted or the transformation of H$_2$S to other sulfur species is underpredicted by the gas phase reactions in the simulation. Overall, the simulation results for H$_2$S show a good agreement with the measurements. Along with the H$_2$S concentrations, the SO$_2$ concentrations are also well predicted because H$_2$S and SO$_2$ are linked through the gas phase reactions.

Figure 12a shows the concentrations of COS for the four air ratios at wall temperature 1000 °C. The right diagram (Figure 12b) depicts the COS concentrations for the three wall temperatures at air ratio 0.8. The measured COS concentrations are between 80 ppm and 400 ppm. The measurements show slightly increasing COS concentrations with decreasing air ratio. There is a good agreement between the measurements and the simulation results for the air ratios 0.8 to 1.0. For air ratio 0.7, the calculated concentration is too low. The simulation results show also increasing COS concentrations with air ratio decreasing from 1.0 to 0.9, but then the calculated concentrations decrease with further decreasing air ratio. This is contradictory to the H$_2$S concentrations (c.f. Figure 10), which are increasing with decreasing air ratio. Figure 12b illustrates the COS concentrations for the different wall temperatures at air ratio 0.8. A trend similar to the H$_2$S concentrations (c.f. Figure 11) can be recognized with elevated concentrations at 1150 °C. The concentration at 1150 °C is slightly higher than for 1000 °C and 1300 °C. The simulation results show the same trend, but the concentrations are between 150 ppm and 200 ppm lower than the measurements. Since the H$_2$S concentrations are well
predicted, the reasons for the deviations in the COS concentrations are probably located in the COS subset of the gas phase reaction mechanism. In general, COS can be consumed either by reaction with the radical pool, by thermal decomposition and by reaction with stable species, such as O$_2$ or H$_2$O. Therefore, the COS reaction subset needs further review.

Figure 11. Measured and simulated H$_2$S and SO$_2$ concentrations at measuring plane 2.
Calculated HCl concentrations are presented for all four air ratios at wall temperature 1300 °C in Figure 13a. There are low concentrations for the air ratios 1.0 and 0.9 and higher concentrations for the air ratios 0.8 and 0.7. The highest measured and calculated concentrations occur at air ratio 0.8. At this point, also the biggest deviation between simulation and measurements appear. The agreement is good for the other three air ratios. Figure 13b depicts the HCl concentration for the three wall temperatures at air ratio 1.0. The calculated HCl concentration is too low for wall temperatures of 1150 °C. The deviation is in the range of 150 ppm. Nevertheless, there is a very good agreement between simulation results and measurements for wall temperatures 1000 °C and 1300 °C. The model is capable of predicting the HCl concentration for high and low temperatures and fails at the middle temperature of 1150 °C.

6. Conclusions

In the present work, 3D CFD simulations of a entrained flow reactor were carried out. A complex mathematical model of pulverized coal combustion including mineral matter transformation and sulfur and chlorine release was presented to predict and assess the high temperature corrosion risk in coal fired furnaces. Measurement data from the entrained flow reactor was considered for evaluating an advanced model for CFD simulations of coal combustion chemistry. The entrained flow reactor data is especially well suited for this purpose since experimental measurements inside the reactor are available and the boundary conditions are well known. This detailed information is crucial for the assessment of the CFD model’s predictive capability for sulfur and chlorine species release and formation. Three different wall temperatures of the reactor and four different air ratios were
investigated, resulting in 12 different cases. The main focus of the work was the modeling of the sulfur and chlorine species behavior during the pulverized coal combustion. The following main conclusions can be drawn from the results reported in this paper.

The concentrations of the main gas species CO₂ and CO show overall a good agreement with the measurements. They lay the foundation for the calculation of trace species concentrations. The applied simulation model delivered good results for H₂S and SO₂ concentrations at substoichiometric conditions. Some of the calculated COS concentrations are lower than the measurements, there is an agreement at other points. Since the H₂S concentrations are well predicted, the reasons for the deviations in the COS concentrations are probably located in the gas phase reaction mechanism, more precisely in the COS subset. The same conclusion can be drawn for the HCl concentrations. Calculated HCl concentrations are too low for some operation points. There is a good agreement at some other operation points.

For future work, ash particle need to be analyzed to evaluate the simulation of the chlorine release. With this information, the chlorine release model and the gas phase mechanism can be adapted. Evaluation of high temperature corrosion risk is still possible because the occurrence of H₂S, COS, and HCl is a sign for possible corrosion attack, even if the exact concentrations may not be known.

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**Abbreviations**

The following abbreviations are used in this manuscript:

Roman
- X mole fraction [-]  
- p pressure [Pa]  
- R ideal gas constant [8314 J/kmol K]  
- T temperature [K]  
- A pre-exponential factor  
- E activation energy [kJ/mol]  
- k rate constant  
- D diffusion constant  
- d particle diameter [m]

Greek
- κ absorption coefficient [1/m]  
- σ scattering coefficient [1/m]  
- ε emissivity [-]  
- Φ phase function [-]

**Abbreviations**

- CFD Computational Fluid Dynamics  
- WSGG weighted sum of gray gases  
- DO Discrete Ordinates  
- RC raw coal  
- Vol-% volume percent
Appendix A

Complete reaction mechanism for the gas phase:

Gas phase reaction mechanism.
Figure A2. Gas phase reaction mechanism.
Gas phase reaction mechanism.

Figure A3.
103. \( \text{HSO} + \text{H} = \text{SH} + \text{OH} \) 
104. \( \text{HSO} + \text{H}_2 = \text{H}_2\text{O} \) 
105. \( \text{HSO} + \text{H}_2\text{S} = \text{H}_2\text{S} + \text{OH} \) 
106. \( \text{HSO} + \text{HO} = \text{HO}_2 + \text{H} \) 
107. \( \text{HSO} + \text{O} = \text{OH} + \text{SO} \) 
108. \( \text{HSO} + \text{OH} + \text{O} = \text{H}_2\text{O} + \text{SO} \) 
109. \( \text{HSO} + \text{O} + \text{H}_2\text{O} = \text{H}_2\text{O}_2 + \text{SO} \) 
110. \( \text{HSO} + \text{H}_2\text{O} = \text{H}_2\text{O} + \text{SO} \) 
111. \( \text{H} + \text{CH}_4 = \text{CH}_3\text{H} + \text{H}_2\) 
112. \( \text{H}_2\text{S} + \text{CH}_3\text{H} = \text{CH}_4 + \text{SH} \) 
113. \( \text{CO} + \text{H}_2\text{O} = \text{H}_2\text{O} + \text{CO} \) 
114. \( \text{CO} + \text{H}_2\text{O} = \text{H}_2\text{O} + \text{CO} \) 
115. \( \text{CO} + \text{H}_2\text{O} = \text{H}_2\text{O} + \text{CO} \) 
116. \( \text{CO} + \text{H}_2\text{O} = \text{H}_2\text{O} + \text{CO} \) 
117. \( \text{CO} + \text{H}_2\text{O} = \text{H}_2\text{O} + \text{CO} \) 
118. \( \text{CO} + \text{H}_2\text{O} = \text{H}_2\text{O} + \text{CO} \) 
119. \( \text{CO} + \text{H}_2\text{O} = \text{H}_2\text{O} + \text{CO} \) 
120. \( \text{CO} + \text{H}_2\text{O} = \text{H}_2\text{O} + \text{CO} \) 
121. \( \text{CO} + \text{H}_2\text{O} = \text{H}_2\text{O} + \text{CO} \) 
122. \( \text{CO} + \text{H}_2\text{O} = \text{H}_2\text{O} + \text{CO} \) 

**UNITS for the preceding reactions (unless otherwise noted):**

- A units mole
- cm
- sec
- K, E units cal/mole

**Figure A4. Gas phase reaction mechanism.**

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